# PREDICTIVE RISK MAPPING OF WATER TABLE DEPTHS IN A BRAZILIAN CERRADO AREA

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## **ABSTRACT:**

Information about hydrological systems and water table fluctuations is important for water management and is needed to assess choices for long-term water management policy. Preferably, uncertainty about the future water table depths is quantified to enable risk assessment. The aim of this study is to map water table depths and quantify the associated uncertainty, in order to indicate extreme scenarios of the hydrological system and areas with potential risks of future water table depths in a Brazilian Cerrado area. To this purpose, a specific type of transfer function-noise (TFN) model, the Predefined Impulse Response Function In Continuous Time (PIRFICT) model is applied. Being the most important driving forces of water table fluctuation, precipitation and potential evapotranspiration are incorporated as exogenous variables into the model. Series of semi-monthly observed water table depths of 37 months length are available from 37 wells in the Jardim river watershed. In the Jardim area almost all natural Cerrado vegetation has been replaced by agricultural crops, some of which are intensively irrigated. Considering that climate conditions of the past 30 years can represent the future climate conditions, statistical characteristics of the future water table dynamics are estimated by stochastic simulation from the calibrated time series models and using 30-year time series on precipitation and evapotranspiration as input. The results are interpolated spatially using Universal Kriging, utilizing a Digital Elevation Model as ancillary information. The resulting probability maps indicate areas with potential risks of water shortage and shallow water table depths for an arbitrary selected date.

# 1. INTRODUCTION

The Brazilian Cerrados extend from the northern margins of the Amazon rain forests to outliers on the southern borders of the country with extensions into Paraguay and Bolivia. Before recent human disturbances, Cerrados probably covered well over 2 million km<sup>2</sup>, equivalent to 23% of the Brazilian territory (Jepson, 2005). The Cerrados are woody savannas, which vary from nearly treeless grasslands to what is almost a woodland of semideciduous trees (Furley, 1999). The structure and physiognomy of the Cerrados reflect climatic and other environmental variables. Total annual rainfall over the central area of Brazilian savannas ranges from 1300 to 1600 mm, concentrated in six to seven months between October and April. The rest of the year is characterized by a pronounced dry season. So, natural vegetation has developed adaptations to the seasonal wet rainfall, acid soils, and aluminium toxicity, and features protective devices against fire. Plants metabolize throughout the year, drawing on soil water reserves, and can withstand short-lived fires.

The expansion of modern agriculture over the past three decades and the establishment of large-scale ranches have affected the characteristic landscape and ecosystem of Cerrados. These areas became Brazil's most important grain belt, facing deforestation rates much higher than in the Amazon rain forest (Oliveira et al., 2005). A complex wood/grass ecosystem was substituted by shallow-rooted monocultures, which are less well adapted to drought. Their need of water supply by irrigation techniques is likely to change the hydrological system (Klink and Moreira, 2000). With irrigation increasing, lowering of the water table occurs, and so risks of

water shortage appear. The preservation of these resources is important because the Cerrado presents itself as an important receiver, keeper and dispersor of water to the tree main hydrographic basins of the country (Amazon basin, Prata basin and São Francisco Basin). The negative impact to its environment might reflect to other Brazilian biomes.

Knowledge about the spatio-temporal dynamics of the water table is important to optimize and balance the interest of economical and ecological land use purposes (Von Asmuth and Knotters, 2004). In hydrology, water table dynamics are modelled in several ways. Many authors refer to transfer function-noise (TFN) models to describe the dynamic relationship between precipitation and the water table depths (Box and Jenkins, 1976; Hipel and McLeod, 1994; Tankersley and Graham, 1994; Van Geer and Zuur, 1997). Basically, these models can be seen as multiple regression methods, where the system is seen as a black box that transforms series of observations on the input (the explanatory variables) into a series of the output variable (the response variable). The parameters of time series models address the temporal variation of the water table depths, while the spatial component can be accessed by regionalizing the outputs using ancillary information related to the physical basis of these models (Knotters and Bierkens, 2000, 2001). This approach can be used to describe the spatio-temporal variation of the water table depths. It is assumed that the spatial differences in water table dynamics are determined by the spatial variation of the system properties, while its temporal variation is driven by the dynamics of the input into the system.

To link the response characteristics of the water table system to the dynamic behaviour of the input, Von Asmuth et al. (2002) presented a method based on the use of a transfer function-noise model in continuous time, the so-called PIRFICT-model. An important advantage of the PIRFICT model as compared to discrete-time TFN-models is that it can deal with input and output series which have different observation frequencies and irregular time intervals. Using a time series model it is possible to simulate over periods without observations, as long as data on explanatory series are available. For instance, long series (say 30 years length) on precipitation and evapotranspiration can be assumed to represent the prevailing climatic conditions (Knotters and Van Walsum, 1997). Alternatively, series generated by climatic models might be applied as inputs to the PIRFICT model. From the simulated realizations, statistical characteristics of future water table dynamics can be calculated, such as mean, standard deviation and limits of prediction.

The aim of this study is to estimate and map the expected water table depths in a watershed located at the Brazilian Cerrados. These estimations are made for a specific date in any future year, given the prevailing hydrological and climatic conditions, in order to support decision making in long term water policy and indicate areas with potential risks of future water shortage and shallow water table depths. In addition, the uncertainty associated with the estimated water table depth is quantified.

### 2. MATERIALS AND METHODS

### 2.1 Study Area

The Jardim River watershed is a representative Cerrado area in the eastern part of the Brazilian Federal District., latitudes  $15^{0}40^{\circ}$ S and  $16^{0}02^{\circ}$ S and longitudes  $47^{0}20^{\circ}$ W and  $47^{0}40^{\circ}$ W. The dry and wet seasons are well defined, with the rainfall concentrated between October and April. During the past years, almost all natural vegetation of the area was replaced by agricultural crops, and the use of irrigation systems has substantially increased in this region during the past years. The main cultivations in the area are soybeans, cotton and corn crops, as well as pasture and horticultural crops. To monitor the water table depths, 37 wells were drilled (Figure 1). The locations were selected purposively, to cover the range of soil types in the area (Lousada, 2005). The water table was observed semi-monthly from October 11, 2003 until October 06, 2006, resulting in series of 50 more or less regularly spaced semimonthly observations, during a period of 1092 days. Series of 33 years length of precipitation and potential evapotranspiration were available from a climate station close to the basin. The period covered is from 1974 until 1996 with a monthly frequency, and from 1996 until March 2007 with a daily frequency. Ancillary information related to geomorphology was derived from a digital elevation model (Figure 1) with 15 meters resolution (Lousada, 2005).

## 2.2 The PIRFICT-model

The behaviour of linear input-output systems can be completely characterized by their impulse response (IR) function (Ziemer et al., 1998; Von Asmuth et al., 2002). The response of water table depth to impulses of precipitation series can be modelled by a transfer function-noise (TFN) model (Box and Jenkins, 1976; Hipel and McLeod, 1994; Von Asmuth and Knotters, 2004). For water table depths, the dynamic relationship between precipitation and water table depth can also be described using

physical mechanistic groundwater flow models. However, by using much less complex TFN models predictions of the water table depth can be obtained which are often as accurate as those obtained by physical mechanistic modelling (Von Asmuth and Knotters, 2004).

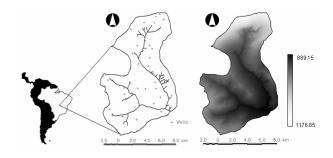


Figure 1. Jardim River watershed and location of the observation wells (+) (left) and the Digital Elevation Model (right).

The basic idea behind TFN modelling is to split the observed series (output) into a sum of transfer components related to known causes (inputs) that influence the temporal variation of the output and an unknown noise component. TFN models are often applied to distinguish between natural and man-induced components of groundwater series (Van Geer and Zuur, 1997). In TFN models one or more deterministic transfer components and a noise component are distinguished. These components are additive. Each transfer component describes the part of the water table depth that can be explained from an input by a linear transformation of a time series of this input. The noise component describes the autoregressive structure of the differences between the observed water table depths and the sum of the transfer components. The input of the noise model is a series of independently and identically distributed disturbances with zero mean, and finite and constant variance, i.e., white noise. Figure 2 shows a scheme of the TFN model.

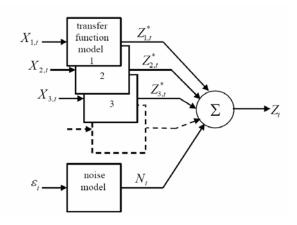


Figure 2. Schematic representation of a transfer function model with an added noise.

In Figure 2,  $X_{1,b}$   $X_{2,b}$   $X_{3,b}$ ... are input variables,  $Z_{1,b}^*$   $Z_{2,b}^*$   $Z_{3,b}^*$ ... are transfer components, and  $\varepsilon_t$  is a stochastic error term with finite and constant variance  $\sigma^2_{\varepsilon}$ .  $N_t$  is the noise component.  $Z_t$  is the output variable. The PIRFICT-model, introduced by Von Asmuth et al. (2002), is a specific type of TFN models and an alternative to discrete-time TFN models. In the PIRFICT-model a block pulse of the input is transformed into an output

series by a continuous-time transfer function. The coefficients of this function do not depend on the observation frequency. The following single input continuous TFN model can be used to model the relationship between water table dynamics and precipitation surplus/deficit. For the simple case of a linear, undisturbed phreatic system that is influenced by precipitation surplus/deficit only (Von Asmuth et al., 2002):

$$h(t) = h^*(t) + d + r(t)$$
 (1)

$$h^*(t) = \int_{-\infty}^t p(\tau)\theta(t-\tau)\partial\tau \tag{2}$$

$$r(t) = \int_{-\tau}^{\tau} \phi(t - \tau) \partial W(\tau)$$
 (3)

where:

h(t) = observed water table depth at time t[T];

 $h^*(t)$  = predicted water table depth at time t attributed to the precipitation surplus/deficit, relative to d[L];

d = level of  $h^*(t)$  without precipitation, or in other words the local drainage level, relative to ground surface [L];

r(t) = residuals series [L];

p(t) = precipitation surplus/deficit intensity at time t [L/T];

 $\theta(t)$  = transfer Impulse Response (IR) function [-];

 $\phi(t)$  = noise IR function [-];

W(t) = continuous Wiener white noise process [L], with properties  $E\{dW(t)\}=0$ ,  $E[\{dW(t)\}^2]=dt$ ,  $E[dW(t_1)dW(t_2)]=0$ ,  $t_1 \neq t_2$ .

The local drainage level d is obtained from the observations as follows:

$$d = \frac{\sum_{i=0}^{N} h(t_i)}{N} - \frac{\sum_{i=0}^{N} h^*(t_i)}{N} - \frac{\sum_{i=0}^{N} r(t_i)}{N}$$
(4)

with N the number of water table depth observations.

TFN models are identified by choosing mathematical functions which describe the IR and the autoregressive structure of the noise. This identification can be done in two ways:

- Iteratively: using correlation structures in the available data and model diagnostics.
- Physically: based on insight into the behaviour of the analyzed system.

Here, the second approach is followed.  $\theta(t)$  is a Pearson type III distribution function (PIII df, Abramowitz and Stegun, 1964). Because of its flexible nature, this function adequately models the response of a broad range of groundwater systems. Under the assumption of linearity, the deterministic part of the water table dynamics is completely determined by the IR function moments. In this case, based on Von Asmuth et al. (2002), the parameters can be defined as:

$$\theta(t) = A \frac{a^n t^{n-1} e^{-at}}{\Gamma(n)}$$

$$\phi(t) = \sqrt{2\alpha\sigma_r^2} e^{-\alpha t}$$
(5)

where A, a, n, are the parameters of the adjusted curve,  $\Gamma(n)$  is the Gamma function and  $\alpha$  determines the decay rate of  $\mathcal{O}(t)$  and  $\sigma^2_r$  is the variance of the residuals.

Equation 5 and its parameters have a physical meaning that is described in Von Asmuth and Knotters (2004). The physical basis of the PIII df lies in the fact that it describes the transfer function of a series of linear reservoirs (Nash, 1958). The parameter n denotes their number and a equals the inverse of the reservoir coefficient normally used. As Knotters and Bierkens (2000) explain, a single linear reservoir (a PIII df with n=1) equals a simple physical model of a one-dimensional soil column, discarding lateral flow and the functioning of the unsaturated zone. The extra parameter A is necessary because in the case of equation 5, where a precipitation and evapotranspiration series are transformed into a water table depths series, the law of conservation of mass does not apply.

The PIII df has shown to be able to model fluctuations of water table closely and comparably to Box-Jenkins TFN models with many more parameters (Von Asmuth et al., 2002). The parameter A is related with the local drainage resistance (the area of the IR function equals the ratio of the mean height of the water table to the mean water table recharge), while Aa is determined by the storage coefficient of the soil and n as the convection and dispersion time of the precipitation through the unsaturated zone. However, care should be taken when interpreting the parameters of the PIII df, or any other time series model for that matter, in physical sense, because of their lumped and empirical nature (Von Asmuth and Knotters, 2004).

After the selection of an IR function that represents the underlying physical process, the available time series have to be transformed to continuous series. First, in order to characterise the variability of precipitation and evaporation, we rely on a simple but effective method to estimate the average precipitation surplus intensity and its annual amplitude. When precipitation surplus data is only available at discrete intervals, the continuous series  $p(\tau)$  cannot be reconstructed exactly, but it can be approximated by assuming that the distribution of  $p(\tau)$  is uniform during the period tpb to tpe (Ziemer et al., 1998). The average level of the precipitation surplus is obtained as:

$$\overline{p} = \frac{\int_{pb}^{pe} p(\tau) \partial \tau}{t_{pe} - t_{ph}} \tag{6}$$

with  $t_{\rm pb}$  and  $t_{\rm pe}$  denoting the start and end of the period over which the meteorological characteristics are calculated. Next, time is split into year Y and the Julian day D, and the precipitation surplus is averaged over Y, which effectively filters out its yearly course:

$$\widetilde{p}(D) = \frac{\sum_{Y_{pb}}^{Y_{pe}} p(Y, D)}{Y_{pe} - Y_{pb}}, \quad 1 \le D \le 365$$
(7)

Because the temperature largely determines the annual evaporation cycle and is more or less harmonic, so is the precipitation surplus, and the annual amplitude can be obtained by matching a sine to the yearly course (Von Asmuth and Knotters, 2004).

Using equation 6, the transfer model (equation 2) can be evaluated using a block response (BR) function  $\Theta(t)$ . The BR function can be obtained convoluting the IR function with a block pulse of precipitation surplus with unit intensity over a period  $\Delta t$ , as follow:

$$\Theta(t) = \int_{t-\Delta t}^{t} \theta(\tau) \partial \tau \tag{8}$$

Because  $\Theta(t)$  is a continuous function,  $h^*(t)$  itself is also continuous, and for every observation of h(t) a sample of the residual series r(t) can be obtained. Next, the noise model (equation 3) is evaluated in order to obtain a series of innovations v(t). Following Von Asmuth et al. (2002), to evaluate the noise model without having to use a Kalman Filter (which is computationally expensive) we will derive a direct relation between the residuals r(t) and the innovations v(t). Consider the series v(t) as the nonequidistantly sampled change in the solution to the stochastic integral describing the residual series:

$$V(t) = \int_{t-\Delta t}^{t} \phi(t-\tau)\partial W(\tau)$$
 (9)

with  $\mathcal{O}(t)$  from equation 5 as the noise IR function, we can rewrite equation 3 as:

$$r(t) = e^{-\alpha \Delta t} r(t - \Delta t) + \int_{-\infty}^{t} \sqrt{2\alpha \sigma_r^2} e^{-\alpha(t-\tau)} \partial W(\tau)$$
 (10)

which is known as an Ornstein-Uhlenbeck process (Uhlenbeck and Ornstein, 1930; Gardiner, 1994). Combining equations 9 and 10, we obtain the innovation series calculated from the available data:

$$v(t) = r(t) - e^{-\alpha \Delta t} r(t - \Delta t)$$
 (11)

Subsequently, an estimative of model parameters set  $\beta$ =(A, a, n,  $\alpha$ ) is made with the aid of a Levenberg-Marquardt algorithm, which numerically minimizes a weighted least squares criterion based on the likelihood function of the noise model. Finally the accuracy and validity of the model are checked using the auto and cross-correlation functions of the innovations, the covariance matrix of the model parameters and the variance of the IR functions. For a complete overview of the PIRFICT model formulation, applications and study cases we refer to Von Asmuth and Maas (2001), Von Asmuth et al. (2002), Von Asmuth and Knotters (2004) and Von Asmuth and Bierkens (2005).

The PIRFICT-model was applied in this study because the model can describe a wide range of response times with differences in sampling frequency between input series and output series. For the Cerrado situation it is particularly interesting because different behaviours of water table can be found even in small catchments. Being the most important driving forces of water table fluctuation, precipitation and evapotranspiration are incorporated as exogenous variables into the model.

## 2.3 Simulating water table depths

Time series models using precipitation surplus/deficit as input variable, calibrated on time series of water table depths with limited years, enable us to simulate series of extensive length (Knotters and Van Walsum, 1997). From extensive series, statistics of WTD can be estimated. These will represent the prevailing hydrological and climatic conditions rather than specific meteorological circumstances during the monitoring period of water table depths.

The simulation of water table depths here presented is based on a time frequency filtering of the PIRFICT-model performed as a convolution in the time frequency domain. This operation considers the shape of the PIII df adjusted from the parameters of each model. Being water table depths the output of a time varying system, the interaction of the precipitation input signal and the system can be regarded as an operation in the time frequency domain between the time frequency expansion of the signal and the time frequency response of the system.

These models contain a dynamic component, describing the dynamic relationship between the input and the output, either physically or empirically. But variation of the water table cannot be completely explained from the precipitation and evapotranspiration series. So, the models must contain a noise component, which describes the part of water table fluctuation that cannot be explained with the used physical concepts or empirically from the input series. The unexplained part (noise component) has to be taken into account in the simulation procedure, since we are interested in statistics of extremes, like the probabilities that critical levels are exceeded. Details about simulation can be found in Hipel and McLeod (1994). Here, the following steps are followed:

After modelling the relationship between precipitation surplus/deficit and water table depths using the PIRFICT-model, series of water table depths are extrapolated to a length of 30 years. It is assumed that the average weather conditions during the last 30 years represent the prevailing climate. As result, deterministic series of predicted water table depths are generated.

- Realizations of the noise process are generated by stochastic simulation and next added to deterministic series resulting in realizations of series of WTD.
   Realizations of the noise process can be generated either by random sampling from a normal distribution with zero mean and residual variance, or by resampling from the fitted residuals.
- From the previous step, N realizations of the stochastic simulation are generated. With a probability density function (PDF) of the distribution of water table depths for each t instant, the statistics representing the prevailing hydrologic conditions can be calculated.

In this study, we applied random sampling from a normal distribution. We decided to calculate statistics about WTD for a t which equals October 1st. This is a reasonable date at which cultivations in the Cerrado region start. The rainy season usually starts around this period. Farmers use to start cultivations just after the first rains of the season.

To enable risk management of water table depths for October 1<sup>st</sup> we calculated two levels of probability. First, a 5 % probability level was considered as a measure for risk of water shortage. With this results, we can say that the area have just 5% probability to have water levels deeper than the values of the resulting map, and 95% higher water levels. The limits established for risks of water shortage at October 1<sup>st</sup> were the deep of the wells, with dry wells characterizing a scenario of water shortage in the area. It can be a problem during the begging of the plants development, affecting water availability and resulting in production losses.

Second, a 95% probability level was considered as a measure for risk of shallow water tables. Shallow water tables can be a problem in the begging of the rainy season because it can make machinery impossible, plow and planting operations. Also it can influence soil conditions, decreasing soil redox potential, increasing pH in acid soils and decreasing in alkaline soils and increasing conductivity and ion exchange reactions. These modifications in the system might influence plant growth, by affecting the availability on toxicity of nutrients, regulating uptake in the rhizosfere. With this results, we can say that the area have just 5% probability to have water levels higher than the values of the resulting map, and 95% deeper water levels. The limits established for risks of shallow water table depths at October 1<sup>st</sup> were 0.5 meters bellow the ground surface.

## 2.4 Regionalizing simulated water table depths

The results of WTD simulations are interpolated spatially using Universal Kriging (Matheron, 1969; Pebesma, 2004). The use of exhaustive information on elevation is interesting, because it can decrease the variance and the uncertainty in the spatial prediction model. Also, when the ancillary information is physically related to target variable it can incorporate physical meaning to the predictions. In our case areas with relatively low elevation and close to drainage devices present relatively shallow water table, whereas in areas with relatively high elevation and far from drainage devices the water table is relatively deep (Furley, 1999).

In order to incorporate DEM as drift (Odeh et al., 1994; Knotters et al., 1995) in the spatial prediction model, this works as follows. Let the simulated water table depths be given as  $z(x_1)$ ,  $z(x_2)$ , ...,  $z(x_n)$ , where  $x_i$  is a (two-dimensional) well location and n is the number of observations (i.e., n=37). At a new, unvisited location  $x_0$  in the area,  $z(x_0)$  is predicted by summing the predicted drift and the interpolated residual (Odeh et al., 1994; Hengl et al., 2004):

$$\hat{z}(x_0) = \hat{m}(x_0) + \hat{e}(x_0) \tag{12}$$

where the drift m is fitted by linear regression analysis, and the residuals e are interpolated using kriging:

$$\hat{z}(x_0) = \sum_{k=0}^{p} \hat{\beta}_k \ q_k(x_0) + \sum_{i=1}^{n} w_i(x_0) \cdot e(x_i);$$

$$q_0(x_0) = 1$$
(13)

Here, the  $\beta_k$  are estimated drift model coefficients,  $q_k(x_0)$  is the kth external explanatory variable (predictor) at location  $x_0$ , p is the number of predictors,  $w_i(x_0)$  are the kriging weights and  $e(x_i)$  are the zero-mean regression residuals. In this case, for WTD, the model was formulated as follows:

$$WTD(x_0) = \beta_0 + \beta_1 \cdot EV(x_0) + e(x_0)$$
 (14)

where EV is the elevation value for each location and e is a zero-mean spatially correlated residual. Its spatial correlation structure is characterized by a semivariogram.

# 2.5 Summary of methods

The methods described in subsections 2.2 to 2.4 are now summarized as follows:

- 1. Calibration of the PIRFICT-model;
- 2. Stochastic simulation of WTD series by using the PIRFICT-model (N=1000) and input series of 30 years length;
- 3. Pick all selected date (October 1<sup>st</sup>) values of WTD generated by stochastic simulation;
- 4. Create a probability distribution function (PDF) of these values;
- 5. Selection of percentile values (5<sup>th</sup> and 95<sup>th</sup>) for WTD;
- 6. Repeat steps 1-5 for all wells;
- 7. Model the spatial structure of the percentile values with geostatistics techniques;

8. Finally, we use these values of WTD to create risk maps of water levels that could be exceeded at the selected date with 5 and 95% probability.

### 3. RESULTS

### 3.1 Time series modelling

Due to spatially varying hydrological conditions, a wide range of calibration results was found for the 37 observed wells. Table 1 summarizes the results of time series modelling.

Table 1. Summary of the statistics of PIRFICT-model calibration

	Min	1 <sup>st</sup> Q	Med	3 <sup>rd</sup> Q	Max	Mean	SD
R <sup>2</sup> adj	57.68	76.16	82.91	88.04	95.45	81.85	9.21
RMSE	0.080	0.350	0.701	0.904	1.886	0.680	0.41
RMSI	0.072	0.313	0.555	0.750	1.433	0.552	0.32

R<sup>2</sup><sub>adj</sub>= Percentage of explained variance; RMSE=Root Mean Squared Error (meters); RMSI=Root Mean Squared Innovation (meters); Min=minimum; 1<sup>st</sup> Q=first quartile; Med=median; 3<sup>rd</sup> O=third quartile; Max=maximum; SD= standard deviation

The percentage of variance accounted for indicate a good fit of the PIRFICT-model to the data. Low percentages might be caused by errors in the data or lack of data, or possibly other inputs that affect the groundwater dynamics are not incorporated into the model (Von Asmuth et al., 2002). The parameters of PIRFICT-model are summarized in Table 2. Some problems with the calibration were diagnosed by checking the impulse response function for each well. After several calibrations, RMSE and RMSI values were the minimum founded for each well.

Table 2. Summary of calibrated parameters of the PIRFICT-model

	Min	1 <sup>st</sup> Q	Med	3 <sup>rd</sup> Q	Max	Mean	SD
A	54.8	770.9	1250	1955	6160	1711.3	705.5
a	0.001	0.004	0.007	0.011	0.162	0.013	0.01
n	0.49	1.05	1.30	1.72	2.86	1.43	0.24
E	-5.85	-0.94	0.88	1.65	2.58	0.24	0.93
α	5.92	22.03	32.24	47.69	95.50	38.60	14.73
IR	0	218	600	900	2200	656.58	570.6
LD	-50.6	-13.9	-7.37	-1.39	0.0	-10.16	11.97

A=drainage resistance (days); a=decay rate (1/days); n=convection time (days); E=reduction factor (-);  $\alpha$ =decay or memory of the white noise process (-); IR=Impulse Response (days); LD=Local Drainage Base (meters); Min=minimum; 1st Q=first quartile; Med=median;  $3^{rd}$  Q=third quartile; Max=maximum; SD= standard deviation

The physical plausibility of the results of a TFN model can be judge, for instance, by checking the IR functions. It is equivalent to the cross-correlation function. We check if the memory of the hydrological system, indicated by the time lag where the IR function approximates to zero, is covered by the monitoring period (De Gruijter et al., 2006). A lack of relationship between the input series and the observed water table depths was found for three wells. The monitoring period apparently was not long enough to characterize the long

memories of the hydrological system in these sites at Jardim

Parameter A is related with the shape of the IR function. Large values of A were calibrated to series at sites where a large fluctuation of the water table level was observed. From a mathematical point of view, these sites have large pulses from the signal of the input in the system. Some attention should be paid to parameter E, the reduction factor of evapotranspiration. This value should be between 0 and 1. For some wells we found estimates of E which are not realistic, like negative values. One reason could be that the climate station, located around 10 km outside the study area, does not represent the meteorological circumstances at all wells locations. Another reason might be in the large temporal variation of land use, which makes these parameters difficult to estimate.

Stochastic simulation with the PIRFICT-model was performed for the 34 wells which remained after inspection of the results. For seven wells the simulation results indicated that the stationary conditions were not met. The distribution functions of the simulated WTD for these wells were bimodal. These wells were excluded from interpolation. Possibly the relatively short length of the water table time series did not completely cover the response time of the hydrological system. For example, the dry years of 2001, 2002 and 2003 might have a long term effect on water table in systems with long memory. The precipitation in these years was 24.4, 41.02 and 33.2% less than the annual average over the last 30 years, respectively. This effect acts different over the basin, due the presence of different geological systems (Lousada, 2005). Continuing monitoring WTD would enable us to clarify these questions (De Gruijter et al., 2006). For the remaining 27 wells, the distribution function of WTD for October 1st was created from the simulated data. The WTDs which are expected to be exceeded with 5% and 95% probability at October 1st were used for spatial interpolation.

# 3.2 Spatial interpolation

The spatial dependence of WTDs that are exceeded with 5% and 95% probability at October 1<sup>st</sup> was modelled by semivariograms (Figure 3). The alternative to use ancillary information on spatial prediction was taken, one time the number of estimation points were sensible reduced. Including elevation as a spatial drift into the geostatistical model caused a decrease in the semivariance.

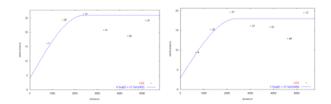


Figure 3. Semivariograms fitted for WTDs that are exceeded at Oct 1<sup>st</sup> with 95% (left) and 5% (right) probability, including a trend that depends on elevation.

The spatial dependence at short distances is poorly estimated because of the small number of observation wells that are fairly uniformly spread across the area. The nugget parameter of the semivariogram reflects the measurement precision of the WTD and the short-distance spatial variation in WTD.

The semivariogram fitted for WTDs that are exceeded with 95% probability, including a trend that depends on elevation, was used on UK estimation. The resulting map shows WTDs that could be exceeded with 95% probability at October 1<sup>st</sup>. The probability to have lower (deeper) values than these in the maps is just 5%. Figure 4 gives these results for October 1<sup>st</sup> of any future year and the corresponding kriging variance.

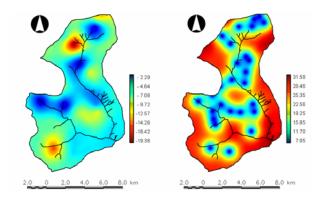


Figure 4. Map of WTDs (meters) that will be exceeded with 95% of probability at Oct 1<sup>st</sup> (right) and the corresponding kriging variance (left).

Figure 5 gives the results of mapping WTD with confident level of 95% including a trend that depends on elevation using and the corresponding kriging variances.

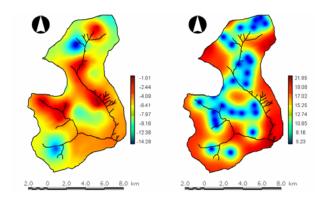


Figure 5. Map of WTDs levels (meters) that will be exceeded with 5% of probability at Oct 1<sup>st</sup> (right) and the corresponding kriging variance (left).

UK resulted in maps with a physical meaning related to the local drainage. These maps were applied to quantify the risks on water management.

We compare the map of WTDs that will be exceeded with 95% of probability with a map of estimated depth of wells as limit for water shortage. It is indicated that no problems with deep water levels will occur at October 1<sup>st</sup>, because no well can be dry with these levels. The same analysis was made for the map of WTDs that will be exceeded with 5% of probability with a 0.5m limit map for shallow water table depths. The risk that these occur at October 1<sup>st</sup> is negligible.

# 3.3 Cross-validation

The results of spatial interpolation were evaluated by cross-validation. Tables 3 and 4 give the results. The SD-values of the observations are much higher than those of predictions,

indicating that the interpolation values are smooth (Eg. 4.86 vs. 1.79 in Table 3 or 4.16 vs. 1.83 in Table 4). The errors can be explained from uncertainty about the calibrated models, a poor relationship between elevation and WTD and from a poor spatial correlation structure in both kriging models.

Table 3. Cross-validation for the spatial interpolation of WTDs which are exceeded with 95% probability at Oct 1<sup>st</sup>.

	Obs.	Pred.	Pred Obs.	Pred. SD	Z-score
Min	-21.36	-14.80	-10.25	3.99	-2.24
1 <sup>st</sup> Q	-11.83	-10.22	-1.99	4.54	-0.42
Med	-8.80	-9.28	1.13	4.77	0.25
$3^{rd} Q$	-7.79	-8.18	2.67	5.04	0.59
Max	-0.72	-5.91	12.35	5.55	2.44
Mean	-9.37	-9.43	-0.05	4.78	-0.006
SD	4.86	1.79	5.49	0.42	1.12

Pred.=predicted; Obs.=observed (meters); Z-score=(Pred.-Obs.) / Kriging variance (-); Min=minimum; 1st Q=first quartile; 3rd Q=third quartile; Max=maximum; SD=standard deviation

Table 4. Cross-validation for the spatial interpolation of WTDs which are exceeded with 5% probability at Oct 1<sup>st</sup>.

	Obs.	Pred.	Pred Obs.	Pred. SD	Z-score
Min	-16.04	-10.73	-7.04	3.36	-1.84
1 <sup>st</sup> Q	-8.42	-6.51	-3.16	3.79	-0.86
Med	-5.62	-5.82	0.78	3.98	0.21
$3^{rd} Q$	-2.62	-4.84	2.24	4.20	0.55
Max	0.14	-1.61	10.23	4.62	2.43
Mean	-5.75	-5.71	0.04	3.99	0.005
SD	4.16	1.83	4.57	0.34	1.12

Pred.=predicted; Obs.=observed (meters); Z-score=(Pred.-Obs.) / Kriging variance (-); Min=minimum; 1st Q=first quartile; 3rd Q=third quartile; Max=maximum; SD=standard deviation

Predictions on Tables 3 and 4 had the mean WTD value respected and indicate small mean interpolation errors (-0.05 and 0.04 meters, respectively). The mean and standard deviation of the *Z*-score had values close to zero and one, respectively, which indicates a good performance of the kriging systems.

## 4. DISCUSSION AND CONCLUSIONS

Time series modelling using the PIRFICT-model was efficient to model a wide range of different responses of the hydrological system presented over the basin. Policy makers can and should use these results to optimize water use and to regulate the competing claims for water resources that often occur between small farmers, big farmers with irrigated crops and water withdrawal for human use. However, the results reflect uncertainties from different sources: uncertainties related to the data (observed WTD, climatic database, DEM), uncertainty associated with time series modelling and with the model of spatial variation.

The quality of the map was restricted by effects of relatively short time series that did not satisfactory characterize long memory systems. The quality of the time series models depends on both sampling frequency and length of series. The quality of the model of spatial structure depends on the number and the configuration of well locations. The use of DEM as ancillary information slightly improved the quality of the final risk maps.

For the chosen date, October 1<sup>st</sup>, there is a negligible risk of water shortage and shallow water table depths that could affect agriculture in some way. The analysis should be extended to other dates and periods that are critical to water supply. The method presented in this study enables this extension.

Given the long memories in the hydrological system of the study area, we recommend to continue monitoring water table depths in order to obtain more reliable results in future.

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